

## Agroforestry creates carbon sinks whilst enhancing the environment in agricultural landscapes in Europe



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### ARTICLE INFO

#### Keywords:

Carbon storage  
Climate change mitigation  
Ecosystem services  
Farmland  
Resource protection  
Spatial deficit analysis

### ABSTRACT

Agroforestry, relative to conventional agriculture, contributes significantly to carbon sequestration, increases a range of regulating ecosystem services, and enhances biodiversity. Using a transdisciplinary approach, we combined scientific and technical knowledge to evaluate nine environmental pressures in terms of ecosystem services in European farmland and assessed the carbon storage potential of suitable agroforestry systems, proposed by regional experts. First, regions with potential environmental pressures were identified with respect to soil health (soil erosion by water and wind, low soil organic carbon), water quality (water pollution by nitrates, salinization by irrigation), areas affected by climate change (rising temperature), and by underprovision in biodiversity (pollination and pest control pressures, loss of soil biodiversity). The maps were overlaid to identify areas where several pressures accumulate. In total, 94.4% of farmlands suffer from at least one environmental pressure, pastures being less affected than arable lands. Regional hotspots were located in north-western France, Denmark, Central Spain, north and south-western Italy, Greece, and eastern Romania. The 10% of the area with the highest number of accumulated pressures were defined as Priority Areas, where the implementation of agroforestry could be particularly effective. In a second step, European agroforestry experts were asked to propose agroforestry practices suitable for the Priority Areas they were familiar with, and identified 64 different systems covering a wide range of practices. These ranged from hedgerows on field boundaries to fast growing

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<https://doi.org/10.1016/j.landusepol.2019.02.025>

Received 19 July 2018; Received in revised form 10 January 2019; Accepted 18 February 2019

Available online 06 March 2019

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coppices or scattered single tree systems. Third, for each proposed system, the carbon storage potential was assessed based on data from the literature and the results were scaled-up to the Priority Areas. As expected, given the wide range of agroforestry practices identified, the carbon sequestration potentials ranged between 0.09 and 7.29 t C ha<sup>-1</sup> a<sup>-1</sup>. Implementing agroforestry on the Priority Areas could lead to a sequestration of 2.1 to 63.9 million t C a<sup>-1</sup> (7.78 and 234.85 million t CO<sub>2eq</sub> a<sup>-1</sup>) depending on the type of agroforestry. This corresponds to between 1.4 and 43.4% of European agricultural greenhouse gas (GHG) emissions. Moreover, promoting agroforestry in the Priority Areas would contribute to mitigate the environmental pressures identified there. We conclude that the strategic and spatially targeted establishment of agroforestry systems could provide an effective means of meeting EU policy objectives on GHG emissions whilst providing a range of other important benefits.

## 1. Introduction

Increased market price volatility and the risks of changing climate are – according to the EU Agricultural Markets Briefs (September 2017) – the biggest challenges European farmers will face in near future (DG Agriculture and Rural Development, 2017). Facing the complex relationship between competitive farming and sustainable production, the current Common Agricultural Policy (CAP, the European framework for agricultural subsidies), supports farmers' income, market measures and rural development (European Commission, 2016). In spite of cross-compliance mechanism and the recently introduced greening measure that links environmental standards to subsidies, the agricultural sector is still one of the prime causes of pressure on natural resources and the environment (EEA, 2017a). To address these environmental problems, the European Commission has issued policies such as the Nitrate Directive (91/676/CEE) in 1991, the Water Framework Directive (Directive 2000/60/EC) in 2000 and the Biodiversity Strategy in 2010 (COM(2011) 244). Nonetheless, major environmental problems persist and are still linked to or caused by intensive agricultural production on the one hand, and by land abandonment on the other (Plieninger et al., 2016). Most recently and in line with the COP21 Paris Agreement (UNFCCC, 2015) the Effort Sharing 2021–2030 (REGULATION (EU) 2018/842) includes agricultural practices, aiming to reduce greenhouse gas (GHG) emissions or balance with an equal amount of GHG sequestration.

In this context, the future CAP for the next funding period after 2020 (CAP2020+) proposes three focal areas: a) “natural” farming, b) sustainable water management and use and c) dealing with climate change (European Commission, 2017a). This will require strategies to manage the above mentioned financial and environmental risks of production, ideas to expand the agricultural product range, and a focus on sustainable farming systems with climate adaptation and mitigation functions (Wezel et al., 2014). Agroforestry, the integrated management of woody elements on croplands or grasslands (European Commission, 2013a), may become part of those strategies because it provides multiple (annual and perennial) products while simultaneously moderating critical environmental emissions and impacts on soil, water, landscapes, and biodiversity (Torralba et al., 2016). In addition, it is highlighted as one of the agricultural practices with the greatest potential for climate change mitigation and adaptation (Aertsens et al., 2013; Hart et al., 2017). For example, agroforestry can enhance the sequestration of carbon in woody biomass and in the soil of cultivated fields (mitigation) (Kim et al., 2016), increase soil organic matter, improve water availability (adaptation to climate aridification) (Murphy, 2015), protect crops, pastures, and livestock from harsh-climate events (adaptation to global warming and increasing wind speed) (Sánchez and McCollin, 2015).

Against this background, our study aimed to evaluate the potential contribution of agroforestry towards achieving zero-GHG emissions agriculture in pursuit of the ambitious Paris Agreement COP21 and CAP targets. Using a transdisciplinary approach including scientific and practical knowledge, the study focused on three key questions: I. Where and to what extent is European agricultural land affected by (multiple)

environmental pressures that could be reduced through agroforestry? II. Which regional types of agroforestry (combinations of various woody plants, crop / animal species and management practices) can be used to reduce these environmental pressures and provide multiple products? and – as an example of an ecosystem service that agroforestry can provide – III. What is the impact of the proposed systems on European climate change targets, in particular on carbon storage and GHG emissions?

## 2. Material and methods

The study was conducted in three main phases: First, the agricultural areas most seriously affected by environmental pressures (“Pressure Areas”) were identified using various spatially explicit datasets on e.g. soil erosion, water pollution, and pollination pressures. In a second step, local agroforestry experts were consulted to propose suitable agroforestry practices for their regions suffering from environmental pressures. Finally, the annual carbon storage impact of the proposed systems was identified and evaluated in the light of European agricultural GHG emissions. In the next subsections, these three main phases are described more in detail, while we address advantages and limitations of the adopted approach, as well as possible improvements, in the Discussion section.

### 2.1. Identification of Priority Areas

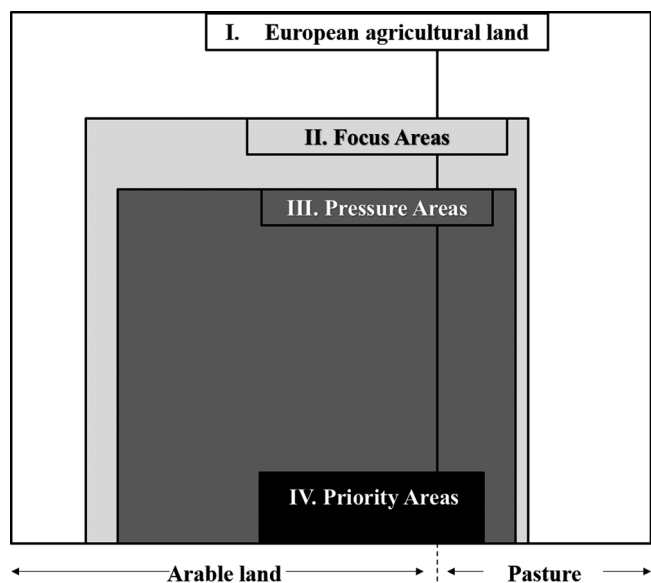
#### 2.1.1. Conceptual approach

Bearing in mind that agroforestry is only one aspect of a diversified agriculture, our focus was on agricultural areas facing combined environmental pressures, in which agroforestry can mitigate several environmental pressures. Fig. 1 illustrates the conceptual background of the Priority Area approach.

The analysis uses the Corine Land Cover 2012 (EEA, 2016) to identify the area of European arable and pasture land. From this farmland layer, the areas of high nature value such as Natura 2000 (EEA, 2015a), High Nature Value Farmland (EEA, 2015b; Paracchini et al., 2008), and the existing agroforestry areas (den Herder et al., 2017) were subtracted. The remaining “Focus Areas” (block II in Fig. 1) were the starting point for the pressure analysis.

#### 2.1.2. Selection of indicators to assess environmental pressures

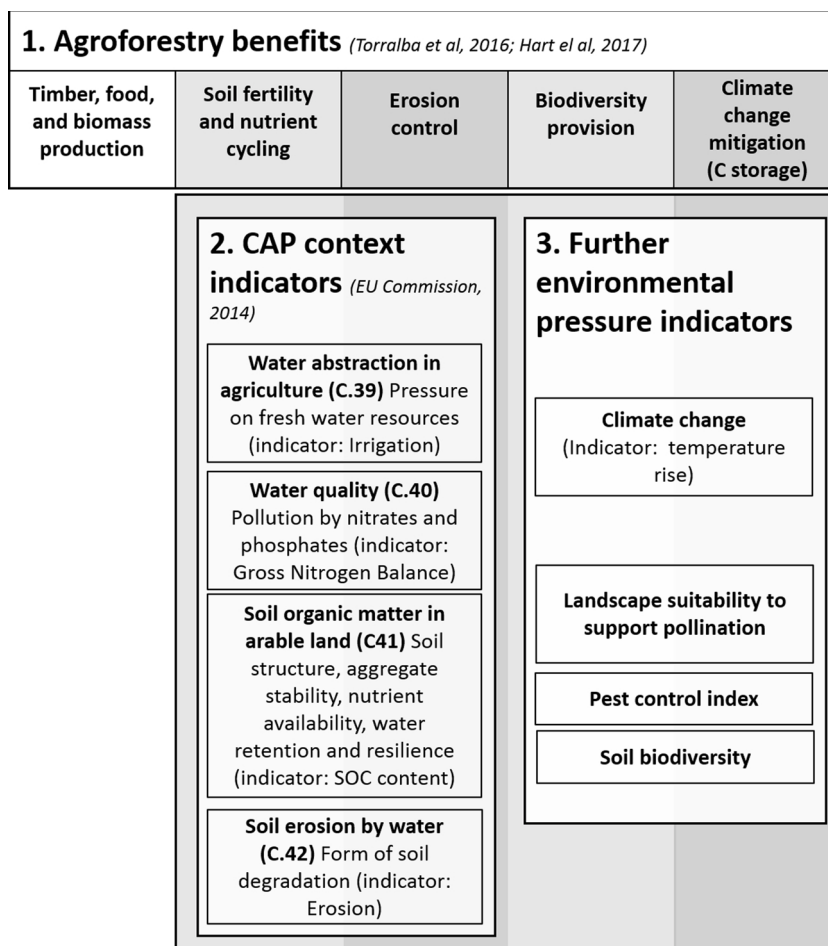
The indicator selection passed three stages as visualised in Fig. 2. First, indicators characterizing benefits provided by agroforestry systems were chosen. Torralba et al. (2016) summarized them into i) timber, food, and biomass production, ii) soil fertility and nutrient cycling, iii) erosion control, iv) biodiversity provision and Hart et al. (2017) completed the list with v) climate change mitigation. Given that continental spatial datasets covering most of the European countries and addressing the indicators in a consistent way were limited, the selection focussed in a second step on the CAP 2014–2020 context indicators. The CAP monitoring and evaluation framework is composed on a set of socio-economic, sectorial, and environmental indicators to reflect the impact and provide (annual) information of the performance



**Fig. 1.** Conceptual approach for the spatially explicit pressure analysis. European agricultural land: Arable and pasture land. Focus Areas: European agricultural land minus nature conservation areas, High Nature Value Farmland, and existing agroforestry land. Pressure Areas: Areas where at least one ecosystem service pressure was mapped. Priority Areas: Areas where environmental pressures accumulate (four out of eight in pasture and five out of nine in arable land).

of the strategy (European Commission, 2017b, 2014). Within this list four context environmental indicators were related to agroforestry benefits. These were i) water abstraction in agriculture (C.39) addressing pressures on available fresh water resources (indicator: e.g. irrigated area), ii) water quality (C.40) dealing with agricultural water pollution by nitrates and phosphates (indicator: e.g. Gross Nitrogen Balance), iii) soil organic matter (SOC) in arable land (C.41) as SOC influences soil structure, aggregate stability, nutrient availability, water retention and resilience (indicator: SOC content), and iv) soil erosion by water (C.42) the most widespread form of soil degradation (indicator: erosion) (European Commission, 2017b). Third, as indicators for climate change mitigation and biodiversity were not addressed within the CAP monitoring, we reviewed the literature and identified relevant datasets. In conclusion, only consistent spatial datasets, which were available with a wide European coverage, were included in the analysis.

Accordingly, environmental pressures related to: i) soil health (soil erosion by wind and water, soil organic carbon), ii) water quality and abstraction (water pollution by nitrates, irrigation), iii) climate change (rising temperature), and iv) biodiversity (pollination and pest control pressures, reduced soil biodiversity) were identified. Individual pressure maps were spatially aggregated and combined into the “Pressure Areas” map showing all regions where one or several environmental pressures occur. To identify the “Priority Areas” for intervention, the sum of pressures per spatial unit (pixel size = 100 m x 100 m) was expressed as an accumulation map or a “heatmap of environmental pressures”.



**Fig. 2.** Selection process of environmental pressure indicators. 1. Assessment of agroforestry benefits, 2. Evaluation of existing CAP context indicators, 3. Review of further environmental pressure indicators for climate change mitigation and biodiversity provision. (SOC: Soil organic carbon).

## 2.2. Pressure Area analysis

### 2.2.1. Soil health pressures

The European water erosion map (Panagos et al., 2015) and the Swiss soil erosion risk map (Prasuhn et al., 2013) together with the European wind erosion map (Borrelli et al., 2017) were used to locate areas with potentially critical loads of soil losses. According to Panagos et al. (2015) a critical threshold is reached if the soil loss is more than 5 t soil ha<sup>-1</sup> a<sup>-1</sup>. The analysis of potential wind erosion was limited to arable land, which is more affected than grassland.

The Soil Organic Carbon (SOC) saturation capacity provided at European level by Lugato et al. (2014a), (2014b) expresses the ratio between actual and potential SOC stocks. Regions with a ratio of less than 0.5 were identified as Pressure Areas, meaning that these soils contain less than half of their SOC storage potential.

### 2.2.2. Water quality pressures

Irrigated fields regardless of whether they were pasture or arable land were included in the pressure analysis. Irrigation maps were provided by the JRC Water Portal (2017) and the Farm Structure Survey (FSS) (Eurostat, 2017a) and expressed the proportion of irrigated land on the total agricultural area. Regions with more than 25% of the agricultural area under irrigation were included as Pressure Area.

The nitrogen surplus, which can lead to both high levels of nitrate leaching and denitrification to gaseous nitrous oxide, was assessed for the European Union using the CAPRI model by Leip et al. (2014). For Switzerland data were obtained from modelled accumulated nitrogen losses (BAFU, 2015). According to the German Ministry of Environment (BMUB, 2017), there is a critical load if the annual nitrogen surplus exceeds 70 kg N ha<sup>-1</sup> a<sup>-1</sup> and this threshold was used to identify areas with high nitrogen surplus.

### 2.2.3. Pressures related to changing climate

Annual mean temperatures from the current climate (1970–2000 WorldClim; Hijmans et al., 2005) and the forecast for 2050 (HadGEM2-ES, Martin et al., 2011) were used to derive the predicted regional

temperature increase up to 2050. According to Hart et al. (2012), agroforestry systems remain robust within an average temperature increase of up to 4 °C. Therefore, all areas with a predicted increase of temperature of more than 2 °C and less than 4 °C were qualified as Pressure Areas where agroforestry could potentially be beneficial.

### 2.2.4. Biodiversity pressures

Soil fauna, microorganisms and biological functions derived from the spatial analysis by Orgiazzi et al. (2016) were used to assess soil biodiversity. The areas identified with “high” and “moderate-high” levels of risk were defined as Pressure Areas.

The pollination assessment was based on the indicator of landscape suitability to support pollinators by Rega et al. (2017). The indicator is a dimensionless score; areas with “very low” and “low” suitability (corresponding to the first two quintiles of the values’ distribution) were defined as Pressure Areas.

The pest control index (Rega et al., 2018) was used as input for the assessment of regions with potential pressures in natural pest control. Again, the indicator is a dimensionless score; areas with “very low” and “low” suitability to support natural pest control, corresponding to the first two quintiles of the values’ distribution, were combined and defined as Pressure Areas.

### 2.2.5. Selection of Priority Areas

Using the thresholds previously mentioned (Table 1), the nine environmental pressures were spatially combined using GIS. In each spatial unit the number of pressures were added together by weighting each indicator equally. Implications, advantages and drawbacks of this methodological approach are addressed in the discussion section (4.1). In the resulting “heatmap”, the 10% of the area with the highest number of pressures were defined as the Priority Area for the implementation of agroforestry. Based on Múcher et al. (2010) the Priority Areas were clustered into seven biogeographical regions: Atlantic; Continental lowlands, Continental hills; Mediterranean lowlands, Mediterranean hills, Mediterranean mountains; and Steppic.

The spatial analysis was performed in ArcGIS10.4 (ESRI, 2016). The

**Table 1**

Spatial datasets with their respective characteristics and the threshold applied to define Pressure Areas (EU28: Austria, Belgium, Bulgaria, Croatia, Cyprus, Czech Republic, Denmark, Estonia, Finland, France, Germany, Greece, Hungary, Italy, Ireland, Latvia, Lithuania, Luxembourg, Malta, the Netherlands, Poland, Portugal, Romania, Spain, Slovakia, Slovenia, Sweden, and the United Kingdom; EU 27: is without Croatia; CH = value for Switzerland).

Indicator	Source	Coverage	Resolution	Threshold
Focus Area	CORINE - Agricultural land	EEA, 2016	all Europe	250 m
	Agroforestry area	den Herder et al., 2017	EU 28, CH	100 m
	High Nature Value Farmland	EEA, 2015b; Paracchini et al., 2008	all Europe (without Greece)	100 m
Soil Pressure Areas	Natura 2000, Ramsar areas	EEA, 2015a	EU 28, CH	
	Soil erosion by water	Panagos et al., 2015; Prasuhn et al., 2013	EU 28, CH	100 m
	Soil erosion by wind	Borrelli et al., 2017	EU 28, CH	500 m
	Soil Organic Carbon (SOC) saturation capacity	Lugato et al., 2014a, 2014b	EU 28	250 m
Water related Pressure Areas	Irrigation	Eurostat, 2017a, 2017b, 2017c; JRC Water Portal, 2017	all Europe	100 m, 1000 m
	Nitrogen surplus	BAFU, 2015; Leip et al., 2014; Hijmans et al., 2005	EU 27, CH (without Cyprus)	1000 m (100 m CH)
Climate Risk Areas	Climate change / Temperature rise		all Europe	
Biodiversity Pressure Areas	Soil biodiversity	Combination of soil fauna, soil microorganisms and soil biological function; Orgiazzi et al., 2016	EU 27	500 m
	Landscape suitability to support pollination	Rega et al., 2017	all Europe (without Cyprus)	100 m
	Pest control index	Rega et al., 2018	all Europe (without Cyprus)	100 m

outcomes were processed in R (R Development Core Team, 2016) with packages plyr (Wickham, 2016), Hmisc (Harrell, 2018), and ggplot2 (Wickham et al., 2016).

### 2.3. Agroforestry recommendations

Potential agroforestry practices, which are: 1) of interest to farmers and the most likely to be adopted by them, 2) the most adapted to mitigate the prominent environmental issues in the region, 3) the most developed in the region and 4) the most suitable to face climate change, were compiled by local experts and the authors for each Priority Area. A total of 20 experts, mainly national delegates of the European Agroforestry Federation (EURAF) or associated researchers, were asked for their contribution.

We used a uniform emailing consisting of an explanation letter, maps of the Priority Area and a structured template. The template was divided into eight questions (see Table 3): i) type of agroforestry (e.g. silvopastoral, silvoarable; hedgerows, coppice, or single trees), ii) title and a short description of the system, iii) tree and hedgerow species, iv) number of trees per hectare or the percentage of woody cover per hectare, v) planting scheme (e.g. lines, scattered) and management system (e.g. year of harvesting / harvesting cycles), vi) crop species and products, vii) tree products, and viii) harvesting year. The outcomes were summarized by biogeographical region.

### 2.4. Assessment of carbon sequestration in biomass

The total biomass production (aboveground wood and root biomass) of the woody elements and the carbon storage potential of the proposed agroforestry systems were assessed based on literature data (see Supplementary material) and from (regional) test sites [units: t biomass ha<sup>-1</sup> a<sup>-1</sup>; t C ha<sup>-1</sup> a<sup>-1</sup>]. Herein the values represented an average potential per year of tree life and did not consider any dynamics of tree growth over time, or other impact factors such as water and nutrient availability, temperature, tree density, etc. Potential minimum and maximum values of carbon storage in biomass (both above- and belowground) of each agroforestry practice for each biogeographic region were extracted separately for pasture and arable land. These values were used for upscaling the results to the “Priority

Area”, assuming that in those regions, the total available farmland would be converted into agroforestry with one of the recommended agroforestry practices.

## 3. Results

### 3.1. Pressure assessment

In EU (EU28 minus Cyprus and Croatia) and Switzerland, the total area of European agricultural land is 1,544,022 km<sup>2</sup> (CLC, EEA, 2016). Subtracting existing agroforestry and nature protection areas, the analysis was then restricted to 1,414,803 km<sup>2</sup> as Focus Area. This area consisted of 1,071,179 km<sup>2</sup> of arable land ( $\pm$  92% of total European arable land) and 343,624 km<sup>2</sup> of pasture (88% of total European pasture).

Fig. 3 gives an overview of the size of the individual “Pressure Areas” in relation to the Focus Area. Soil loss risks over 5 t soil ha<sup>-1</sup> a<sup>-1</sup> from water erosion were identified on 11.9% of the arable area and 9.5% of the pasture. Areas suffering from an annual loss greater than 5 t soil ha<sup>-1</sup> a<sup>-1</sup> by wind erosion were relatively small (1.5%), whereas a low SOC saturation capacity was present on 58.7% of arable lands and on 12.8% of pastures. In total, 8.4% of the arable areas and 1% of the pastures had irrigation levels greater than 25%. High nitrogen pollution risk was mapped on 20.6% of arable lands and on 34.5% of the pastures. Around 63.0% of arable lands and 53.6% of pastures were located in regions where temperature is expected to rise between 2 and 4 °C by 2050 according to the HadGEM2-ES forecast scenario. Pressures in biodiversity and resulting potential underprovision of ecosystem services are widely spread all over European agricultural land. In total, 66.4% of the arable lands and 36.8% of pastures in the Focus Area were predicted to have low or very low natural pest control potential, whilst 41.8% of the arable areas and 21.0% of pastures were predicted to be not suitable for supporting pollinators. Potential soil biodiversity pressures were mapped on 11.5% of arable lands and on 18.7% of pastures.

By combining the nine individual pressure maps, we created a heatmap for environmental pressures (Fig. 4a).

For the total Pressure Area, a lower proportion of pasture areas were identified than of arable lands. Only 4% of the arable lands in the Focus

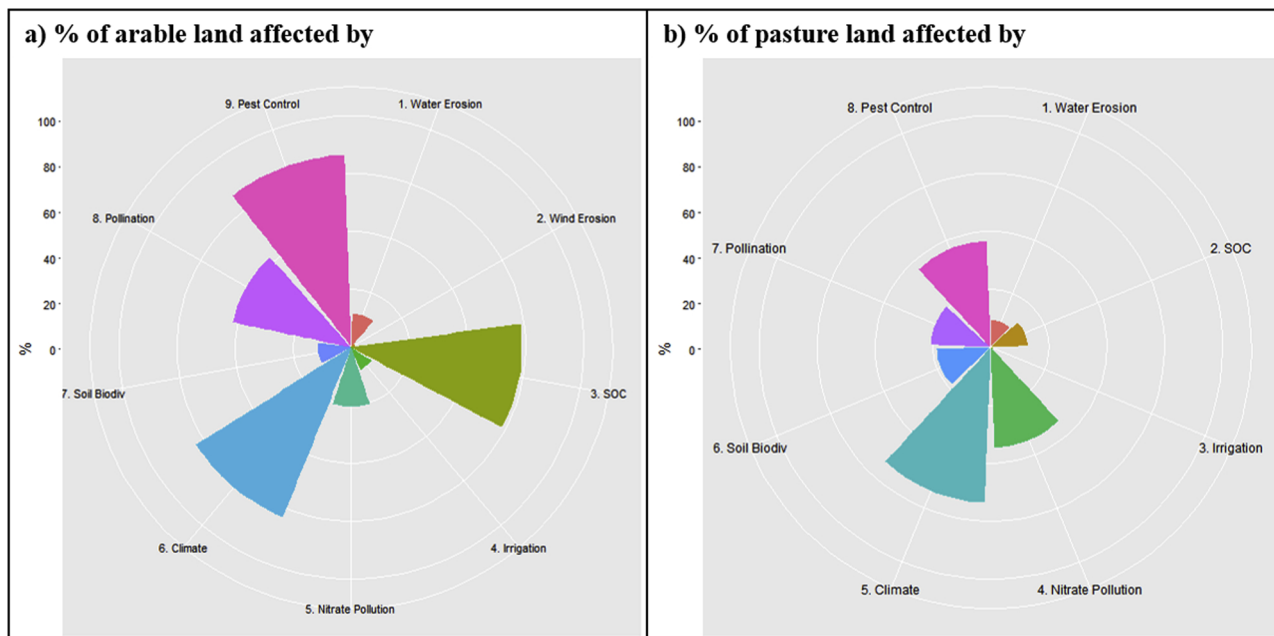


Fig. 3. The proportion of a) the arable land and b) the pasture affected by each of the environmental pressures across the selected Focus Areas. Wind erosion was only considered in arable areas. (SOC: Soil organic carbon).

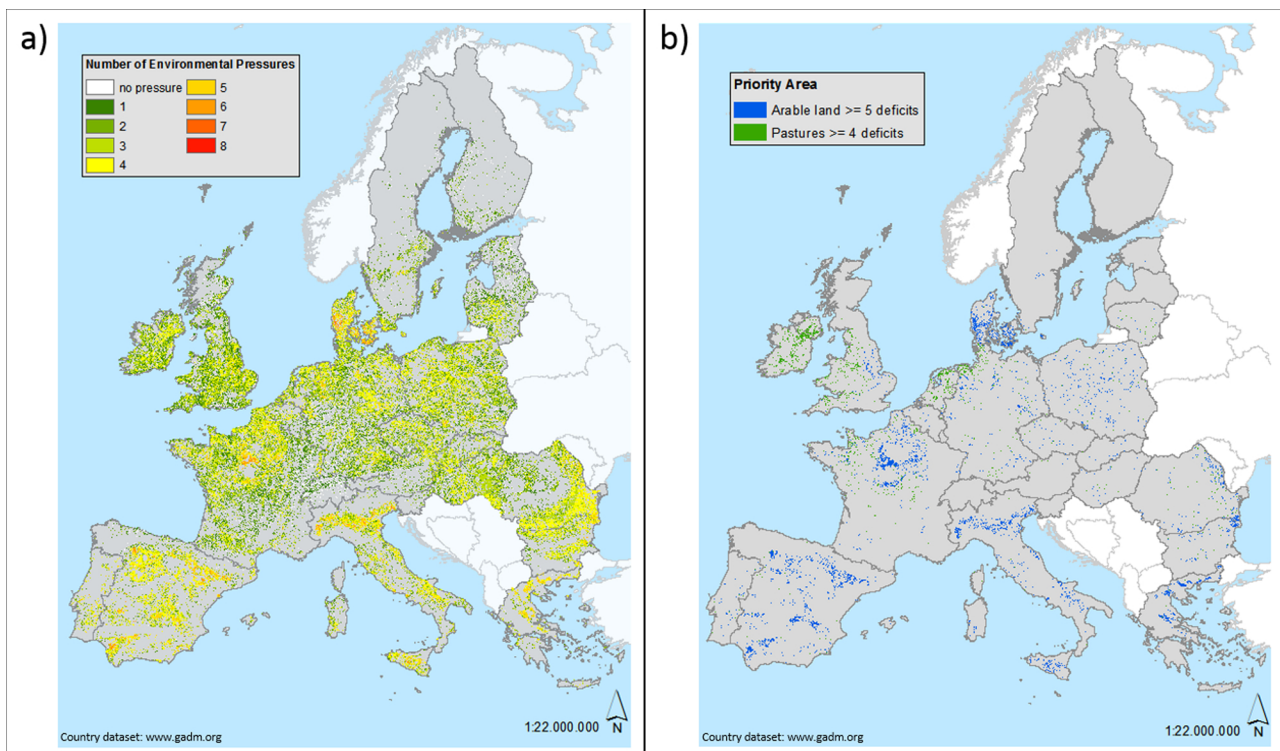


Fig. 4. a) Heatmap for the number of environmental pressures and b) Priority Areas (arable areas with more than five pressure indicators; and pasture areas with more than four pressure indicators).

**Table 2**  
Summary of the Priority Areas by country divided into biogeographical regions based on the landscape classification by Múcher et al. (2010).

Biogeographical Region	Country	Arable land [km <sup>2</sup> ]	Pasture [km <sup>2</sup> ]	Total [km <sup>2</sup> ]	Share of total agricultural land [%]	
Atlantic	<b>Total</b>	<b>29,611</b>	<b>29,088</b>	<b>58,698</b>	<b>9.74</b>	
	Denmark	498	3,223	3,721	20.19	
	France	16,156	6,151	22,308	10.7	
	Germany	6,366	102	6,468	9.78	
	Ireland	6	7,133	7,139	17.12	
	Netherlands	2,624	3,030	5,654	32.96	
	UK	2,600	8,719	11,319	8.43	
	others	1,361	730	2,090	1.8	
	Continental	<b>Lowlands</b>	<b>7,644</b>	<b>1,259</b>	<b>8,903</b>	<b>6.24</b>
		Denmark	3,607	21	3,628	38.82
Germany		1,660	809	2,469	5.22	
Poland		1,296	106	1,402	3.76	
Others		1,081	322	1,403	2.88	
<b>Hills</b>		<b>13,906</b>	<b>4,360</b>	<b>18,265</b>	<b>4.11</b>	
Bulgaria		2,116	537	2,654	7.03	
Germany		1,905	1,473	3,377	3.88	
Poland		6,379	439	6,818	5.73	
Romania		2,054	1,078	3,132	4.87	
Mediterranean	<b>Lowlands</b>	<b>12,399</b>	<b>156</b>	<b>12,555</b>	<b>22.52</b>	
	Greece	3,020	42	3,063	38.28	
	Italy	7,990	39	8,029	21.15	
	Spain	1,220	50	1,270	22.36	
	Others	169	25	193	4.7	
	<b>Hills</b>	<b>20,226</b>	<b>650</b>	<b>20,876</b>	<b>15.53</b>	
	Greece	2,340	117	2,457	22.04	
	Italy	6,985	83	7,069	15.64	
	Spain	9,676	227	9,903	25.02	
	Others	1,225	223	1,448	3.77	
Mountains	<b>Total</b>	<b>12,858</b>	<b>628</b>	<b>13,486</b>	<b>10.96</b>	
	Italy	1,071	78	1,149	10.66	
	Spain	11,176	429	11,606	12.34	
	Others	611	120	732	4.02	
Steppic	<b>Total</b>	<b>2,948</b>	<b>1,026</b>	<b>3,974</b>	<b>11.54</b>	
<b>Total</b>		<b>99,592</b>	<b>37,166</b>	<b>136,758</b>	<b>8.87</b>	

**Table 3**  
Agroforestry practices for arable lands and pastures in the European biogeographical regions (an extract of the practices with the lowest, medium, and the highest carbon sequestration potential are shown. See Supplementary material for the complete list and references). (SRC: Short rotation coppice)

Biogeographical region	Agroforestry type	Title	Tree / hedgerow species	Trees [trees ha <sup>-1</sup> ], hedgerow [m ha <sup>-1</sup> ] or wood cover [% ha <sup>-1</sup> ]	Management system	Crop species and products	Tree products	Year of tree harvesting	Carbon sequestration [t C ha <sup>-1</sup> a <sup>-1</sup> ]
Atlantic pasture	Silvopastoral, coppice	Agroforestry for ruminants in France	Pear ( <i>Pyrus</i> spp), honey locust ( <i>Gleditsia triacanthos</i> ), service tree ( <i>Sorbus domestica</i> ), white mulberry ( <i>Morus alba</i> ), Italian alder ( <i>Alnus cordata</i> ), goat willow ( <i>Salix caprea</i> ), field elm ( <i>Ulmus minor</i> ), black locust ( <i>Robinia pseudoacacia</i> ), grey alder ( <i>Alnus incana</i> )	(single – 2 m, double – 6 m, triple – 10 m), 4 m for trees, 1.3 m coppices x 20 m, (11% woody cover)	Single, double, or triple lines	Grazing, hay, silage	Fodder-trees, woodchips	5–8	0.16-0.48
Atlantic pasture	Silvopastoral, single trees	Traditional orchard	Fruit trees (apple – <i>Malus domestica</i> , pear - <i>Pyrus</i> spp, plum - <i>Prunus domestica</i> )	80 trees ha <sup>-1</sup>	Lines	Grazing, hay, silage	Fruits (woodchips)	60	1.23
Atlantic pasture	Silvopastoral, single trees	High stem timber trees	Poplar ( <i>Populus</i> spp)	400 trees ha <sup>-1</sup> , After 15-20 years: 120-150 trees ha <sup>-1</sup>	Lines	Grazing, hay, silage	Timber	First cut: 15-20 harvest:25-30	2.78-6.35
Atlantic arable	Silvoarable, hedgerows	Productive boundary hedgerow	Mixed hedgerow species: hawthorn ( <i>Crataegus</i> spp), blackthorn ( <i>Prunus spinosa</i> ), field maple ( <i>Acer campestre</i> ), hazel ( <i>Corylus avellane</i> )	0.03% ha <sup>-1</sup>	Boundary hedgerow	Crop rotation with cereals (wheat, barley, oats), potatoes, squash, organic fertility building ley	Woodchips	Every 15	0.1-0.45
Atlantic arable	Silvoarable, coppice	Alley cropping – Short Rotation Coppice (SRC)	Willow ( <i>Salix viminalis</i> ), hazel ( <i>Corylus avellana</i> )	1000-1300 trees ha <sup>-1</sup> (24% ha <sup>-1</sup> )	Twin rows with 10-15 m wide crop alley	Cereals (wheat, barley, oats), potatoes, squash, organic fertility building ley	Woodchips	Every 2 for willow, every 5 for hazel	0.36-1.05
Atlantic arable	Silvoarable, single trees	High stem timber trees	Walnuts ( <i>Juglans regia</i> ), maples ( <i>Acer</i> spp), wild cherry ( <i>Prunus avium</i> ), checker tree ( <i>Sorbus torminalis</i> ), service tree ( <i>Sorbus domestica</i> ), apple ( <i>Malus domestica</i> ), pear ( <i>Pyrus</i> spp)	28-110 trees ha <sup>-1</sup> , (26-50 m between rows)	Lines	Grazing, hay, silage	Timber	60	Walnut: 0.32-2.75, cherry: 0.19-1.4
Continental hills pasture	Silvopastoral, single trees	Wooded grassland	Fruit trees: cherry ( <i>Prunus avium</i> ), walnut ( <i>Juglans regia</i> ), apple ( <i>Malus domestica</i> ), etc.	60 trees ha <sup>-1</sup>	Lines	Grazing, hay, silage	Fruits	70-90	Cherry: 0.41-0.76, apple: 0.93-1.43, walnut: 0.86-1.16 Poplar: 0.44-1.41
Continental lowlands pasture	Silvopastoral, coppice	Agroforestry for free-range pig production	Poplar ( <i>Populus</i> spp), willow ( <i>Salix</i> spp), various fruit trees	10-40% ha <sup>-1</sup> (2.5 × 3.5 m)	SRC lines	Grazing, hay, silage	Woodchips, fodder-trees	5-8	
Continental hills pasture	Silvopastoral, single trees	High nature and cultural value wood pastures and wooded grasslands	Sessile oak ( <i>Quercus petraea</i> ), beech ( <i>Fagus sylvatica</i> ), hornbeam ( <i>Carpinus betulus</i> ), wild fruit trees, mixed poplar ( <i>Populus</i> spp), willow ( <i>Salix</i> spp)	50-300 trees ha <sup>-1</sup> (10-50% ha <sup>-1</sup> )	Scattered	Grazing, hay, silage	Acoms, fruits, timber, (fodder-trees)	Trees not harvested	Oak: 0.71-2.83, beech: 0.59-2.34, hornbeam: 0.38-1.55
Continental lowlands arable	Silvoarable, coppice	Alley cropping	Poplar ( <i>Populus</i> spp); Mixed hedgerow species: willow ( <i>Salix</i> spp), hornbeam ( <i>Carpinus betulus</i> ), common ash ( <i>Fraxinus excelsior</i> ), common birch ( <i>Betula pendula</i> ), black locust ( <i>Robinia pseudoacacia</i> )	Rows A, B, and C: 10,000 trees ha <sup>-1</sup> , Rows D, E, F, and G: 2222 trees ha <sup>-1</sup> , (10% ha <sup>-1</sup> ).	Single and twin rows with 48, 96, and 144 m wide crop alleys.	Crop rotation (wheat, maize, oilseed rape, barley)	Woodchips	Rows A, B, and C: every 3-5. Rows D, E, F, and G: every 8-10	0.15 - 0.44
Continental hills arable	Silvoarable, single trees	Orchard with vegetables or fruits (strawberries)	Fruit trees: cherry ( <i>Prunus avium</i> ), walnut ( <i>Juglans regia</i> ), apple ( <i>Malus domestica</i> ), etc	60 trees ha <sup>-1</sup>	Lines	Vegetable, berries (strawberries)	Fruits, timber	70-90	Cherry: 0.41-0.76, apple: 0.93-1.43, walnut: 0.86 -1.16
Continental hills arable	Silvoarable, single trees	Non-native, energy tree with Alfalfa	Paulownia ( <i>Paulownia tomentosa</i> )	126 trees ha <sup>-1</sup> (18 m x 5 m)	Lines	Triticale, alfalfa	Timber	10-12	3.77

(continued on next page)

Table 3 (continued)

Biogeographical region	Agroforestry type	Title	Tree / hedgerow species	Trees [trees ha <sup>-1</sup> ], hedgerow [m ha <sup>-1</sup> ] or wood cover [% ha <sup>-1</sup> ]	Management system	Crop species and products	Tree products	Year of tree harvesting	Carbon sequestration [t C ha <sup>-1</sup> a <sup>-1</sup> ]
Mediterranean hills pasture	Silvopastoral, single trees	Dehesa	Holm oak ( <i>Quercus ilex</i> )	25–50 trees ha <sup>-1</sup>	Scattered	Grazing	Acorns, fodder-trees	Trees not harvested	0.09–0.16
Mediterranean hills pasture	Silvopastoral, single trees	Grazed cork oak plantation	Cork oak ( <i>Quercus suber</i> )	113 trees ha <sup>-1</sup> , after 20 years: 50 trees ha <sup>-1</sup>	Lines	Grazing	Cork, timber	80	0.34–1.29
Mediterranean hills/mountains pasture	Silvopastoral, single trees	Grazed fruit plantations	Olive ( <i>Olea europaea</i> ), almond ( <i>Prunus dulcis</i> )	250 trees ha <sup>-1</sup>	Lines	Grazing, legume rich mix (annual self seeding species)	Fruits, oil, nuts	Annual prunings, trees not harvested	Olive: 1.97, almond: 1.36
Mediterranean lowlands arable	Silvoarable, single trees	High stem timber trees	Pedunculate oak ( <i>Quercus robur</i> )	57 trees ha <sup>-1</sup>	Lines	Cereals	Timber	35	0.11–0.26
Mediterranean hills arable	Silvoarable, single trees	Fruit tree alley	Olive ( <i>Olea europaea</i> )	200–400 trees ha <sup>-1</sup>	Lines or scattered	Wild asparagus	Oil, forage	Annual prunings, trees not harvested	1.57–3.14
Mediterranean mountains arable	Silvoarable, single trees	High stem timber trees	Poplar ( <i>Populus</i> spp)	200 trees ha <sup>-1</sup>	Lines	Crop rotation wheat, oilseed rape, chickpeas	Timber	15	5.76 - 7.29
Steppic arable	Silvoarable, single trees	High stem forest trees	Poplar ( <i>Populus</i> spp), willow ( <i>Salix</i> spp.), black locust ( <i>Robinia pseudoacacia</i> ), pedunculate oak ( <i>Quercus robur</i> ), plain common and black walnut ( <i>Juglans nigra</i> ), common ash ( <i>Fraxinus excelsior</i> ), red oak ( <i>Quercus subra</i> ), lime ( <i>Tilia</i> sp.), Grayish oak ( <i>Quercus pedunculiflora</i> ), field maple ( <i>Acer campestre</i> ), lime ( <i>Tilia</i> sp.), hawthorn ( <i>Crataegus</i> sp), <i>Rosa</i> sp, blackthorn ( <i>Prunus spinosa</i> )	60–70 trees ha <sup>-1</sup>	Lines	Vegetables	Timber	70–90	Poplar: 1.72–2.85, oak: 0.32–1.2, walnut: 1.31
Steppic, arable	Silvoarable, single trees	Mixed timber and wild fruit species plantation	Poplar ( <i>Populus</i> spp)	100 trees ha <sup>-1</sup>	Lines	Vegetables	Fruits, fodder trees, timber	Harvesting depends on species estimated from 25–120.	Oak : 1.59, tilia: 1.32
Steppic, arable	Silvoarable, single trees	Poplar plantation	Poplar ( <i>Populus</i> spp)	100 trees ha <sup>-1</sup>	Lines	Sunflower, cabbage, corn, pepper and eggplant, water-melon and squash, cauliflower, wheat, beans	Timber	35	2.88–4.76

Areas had no pressures, while in pasture it was around 12%. More than half of the pasture areas had less than three pressures, while 35% of arable area were affected by more than four pressures, and 9% had more than five pressures. Whilst we defined the Priority Areas as arable lands with more than five pressures, we set the threshold to only four pressures for pasture, as we evaluated only eight pasture pressure indicators (excluding soil erosion by wind). Together, they represent the worst 10% of the Pressure Area (Fig. 4b). These combined Priority Areas for arable and pasture land amounted to 136,758 km<sup>2</sup>, which corresponds to about 8.9% of the total European agricultural land. Table 2 gives an overview of the Priority Areas according to country and biogeographical region.

### 3.2. Potential agroforestry practices

In total, 64 agroforestry practices were proposed by the authors and local experts. They cover a wide range of practices from hedgerow systems on field boundaries to fast growing coppices or scattered single tree systems. Table 3 lists, for each biogeographical region, the proposed system with the lowest, medium, and highest carbon sequestration potential. In line with the largest Pressure Areas, the highest number of agroforestry practices was proposed for Atlantic regions (14 silvopastoral and 9 silvoarable practices) followed by Mediterranean arable lands. The complete list can be found in Supplementary material.

### 3.3. Carbon storage potential

For each system the annual carbon storage potential of the woody elements (including roots) was identified using data from the literature and in each geographical region, the minimum and maximum storage potential were determined. The wide range of practices selected corresponded to a wide range of carbon storage potentials, between 0.09 and 7.29 t C ha<sup>-1</sup> a<sup>-1</sup>. In Table 4 these data were upscaled to the entire Priority Area of each biogeographical region. Overall, implementing the proposed agroforestry practices in the Priority Areas could mitigate between 2.1 and 63.9 million t C a<sup>-1</sup> depending on the systems chosen, which is between 7.7 and 234.8 million t CO<sub>2eq</sub> a<sup>-1</sup>.

In 2015, the 28 members of the European Union (EU28) together with Switzerland emitted 4,504.9 million t of greenhouse gases (million t CO<sub>2eq</sub>), with agriculture contributing 12% (~540 million t CO<sub>2eq</sub>; Eurostat, 2017b). Converting the conventionally used farmland in the Priority Area (which was about 8.9% of total agricultural land) to agroforestry could therefore capture between 1.4 and 43.4% of the European agricultural GHG emissions.

## 4. Discussion

This research investigated three questions: I) Where and to what extent is European agricultural land affected by (multiple)

environmental pressures? II) Which regional types of agroforestry can be used to reduce environmental pressures? and III) What is the potential contribution of the proposed systems to the European zero-emission agriculture climate targets?

### 4.1. European environmental Pressure Areas

In response to the first question, several environmental pressures that can be mitigated by establishing agroforestry practices were selected. According to Alam et al. (2014) and Torralba et al. (2016) these include soil conservation, the improvement of water quality, nutrient retention, climate regulation, and enhanced biodiversity. We investigated nine environmental pressures and mapped their occurrence in European agricultural land, based on existing spatially explicit databases at a continental European scale. The best available data were used, although it should be noted that differences in scales (100–1000 m pixel size), time periods (2006–2017) and models (e.g. modelled soil losses in EU vs. soil erosion risk map in Switzerland) existed that might result in spatial inaccuracies (Schulp et al., 2014). However, as other authors have pointed out, this is an intrinsic limit of all pan-European, spatially explicit studies: “as fully harmonized data on the different aspects are not available, the possible bias from inconsistencies between the different data layers is unavoidable” (Malek and Verburg, 2017). All the datasets used, required some degree of modelling and the maps therefore show predicted rather than measured environmental pressures. Moreover, not all the existing environmental problems in agricultural areas could be addressed. Methane emissions, ammonia emissions, and zoonoses contamination, for example, were not included in the analysis presented here. In addition, biodiversity aspects in terms of quality and diversity (Zhang et al., 2007), the amenity value of the landscape, and natural hazards, such as avalanches, floods, droughts, and landslides (EEA, 2017b) were not considered.

Recommendations from the literature were used to define the thresholds for delimiting the Pressure Areas. The definition of thresholds is always arbitrary to some extent: different thresholds exist and modifying these or using different models would affect the size and spatial location of the Pressure Areas. For erosion, we used 5 t soil ha<sup>-1</sup> a<sup>-1</sup> as a threshold for erosion caused by water and erosion caused by wind, whereas for example, adopting a “tolerable” soil erosion rate of 0.3 to 1.4 t soil ha<sup>-1</sup> a<sup>-1</sup> as recommended by Verheijen et al. (2009) would strongly have increased the Pressure Area. The 5 t soil ha<sup>-1</sup> a<sup>-1</sup> threshold was uniformly used for the whole of Europe. However, soil erosion threshold values could also be defined by the nature of the soils in a particular area, depending for example, on soil quality and depth, with lower quality and shallower soils given lower thresholds to reflect their already precarious state and the relative importance of conserving what remains.

Surplus regions for nitrogen have also been defined in different

**Table 4**

Potential carbon sequestration in the whole Priority Area using minimum and maximum carbon storage potential of agroforestry practices proposed for each biogeographical region.

Biogeographical region	Minimum carbon storage potential					Maximum carbon storage potential				
	[t C km <sup>-2</sup> a <sup>-1</sup> ]		Priority Area [t C a <sup>-1</sup> ]			[t C km <sup>-2</sup> a <sup>-1</sup> ]		Priority Area [t C a <sup>-1</sup> ]		
	Arable land	Pasture	Arable land	Pasture	Total	Arable land	Pasture	Arable land	Pasture	Total
Atlantic	10	16	296,109	465,401	761,510	275	635	8,142,998	18,470,618	26,613,616
Continental lowlands	15	44	114,660	55,396	170,056	159	141	1,215,401	177,518	1,392,919
Continental hills	27	38	375,461	165,661	541,122	377	283	5,242,545	1,233,741	6,476,286
Mediterranean lowlands	11	9	136,390	1,400	137,790	600	197	7,439,447	30,654	7,470,101
Mediterranean hills	11	9	222,488	5,850	228,338	530	197	10,719,872	128,053	10,847,925
Mediterranean mountains	11	9	141,441	5,650	147,092	729	197	9,373,711	123,676	9,497,387
Steppic hills	32	38	94,322	39,003	133,325	476	283	1,403,039	290,467	1,693,506
Total			<b>1,380,871</b>	<b>738,362</b>	<b>2,119,233</b>			<b>43,537,013</b>	<b>20,454,727</b>	<b>63,991,740</b>

ways by the European states. The Nitrate Directive (91/676/CEE) limits the nitrate content in ground and drinking waters to  $50 \text{ mg NO}_3 \text{ l}^{-1}$ , and uses this limit for national governments to identify Nitrate Vulnerable Zones (NVZ). In an earlier study on arable target regions for agroforestry implementation, based on soil erosion risk and NVZs, Reisner et al. (2007) identified 51.6% of the European arable land as Pressure Area. Yet the delimitation of NVZs was partly also a political process. In some countries they are limited to areas where the nitrate content in groundwater regularly exceeded the  $50 \text{ mg NO}_3 \text{ l}^{-1}$  threshold. In other countries, entire territories or regions were designated where special actions for nitrate reduction are compulsory for farmers (European Commission, 2013b). For example, almost the entire territory of Germany is labelled as NVZ. To allow for a spatially more differentiated analysis, we opted to locate areas with modelled annual nitrogen surplus above  $70 \text{ kg N ha}^{-1}$ . Together, they accounted for 22% of arable lands and 36% of pastures, which is substantially lower than the 51.6% of European arable land identified by Reisner et al. (2007) as Pressure Area for nitrate emissions.

The most prominent pressure in terms of area affected was the impact of rising temperature and climate change. This is in line with Olesen et al. (2012) and Schauburger et al. (2017) who modelled effects of climate change on crop development and yields. They found an earlier start to the growing and flowering period followed by enhanced transpiration in combination with water stress resulted in a reduction of maize yield of up to 6% for each day with temperatures over  $30^\circ\text{C}$ . In fact, already during the summer of 2017 the potential impact of climate change was revealed by drought and heat waves, which impeded cereal production in various parts of Europe, mainly in southern and central Europe (JRC, 2017). However, by contrast, Knox et al. (2016) predicted positive effects of between 14–18% on the yields of wheat, maize, sugar beet, and potato by 2050 in Northern Europe.

To identify Priority Areas, we accumulated all indicators. This simple addition implied assigning the same weight to all the environmental pressures addressed, and not considering the magnitude of each pressure and its relevance for the local context. For instance, soil erosion could be more damaging for agricultural practices than pests in a particular region or vice versa. However, a more sophisticated approach incorporating these two aspects, would have introduced a further level of arbitrariness in the study, in relation to the assignment of different weights. The approach used here has the advantage of being straightforward and immediate to understand and interpret for decision-makers. Indeed, our methods and results are in line with other pan-European studies, e.g. Mouchet et al. (2017) and Maes et al. (2015), that both analysed the ecosystem service provision of European landscapes. Mouchet et al. (2017) aggregated bundles of ecosystem services and found a longitudinal gradient of decreasing land use intensity from France to Romania. Maes et al. (2015) assessed the quantity of green infrastructure that maintained regulating ecosystem services and showed that regions with intensive agricultural production (arable and livestock) generally had lower levels of regulating ecosystem services provision. Both studies referred to the sum of all assessed indicators. The similarity among the three studies for the spatial output gives confidence to the overall outcomes of this study.

#### 4.2. Potential agroforestry practices and ecosystem service provision

To address the second research question, the collection of agroforestry practices, we hypothesized that agroforestry could mitigate the environmental pressures identified and that for each region, suitable practices could be proposed. Although agroforestry provides multiple ecosystem services (Torralba et al., 2016), there is a general lack of uptake by farmers (Rois-Díaz et al., 2018). Therefore, instead of trying to propagate the most suitable agroforestry for a particular pressure area and environmental pressure, we argue that the highest impact could be achieved by proposing an array of agroforestry practices that are locally adapted and attractive for farmers. This was how the experts

selected the proposed practices. The suitable combination of tree and crop species is highly dependent on soil, water, and climate conditions at specific locations. For this reason, we have provided only a list of examples of agroforestry practices. The composition, implementation, and management of the agroforestry systems needs to be discussed with regional agroforestry experts and developed in partnership with the farmers themselves<sup>1</sup>.

For soil conservation, silvoarable alley cropping systems have been evaluated in earlier studies. Palma et al. (2007) and Reisner et al. (2007) estimated that their introduction on eight million hectares of arable land subject to water induced erosion risks would reduce soil erosion in those areas by 65%. Similar findings were provided by Ceballos and Schnabel (1998) and McIvor et al. (2014), who analysed how agroforestry can contribute to soil protection and preservation. Hedgerow systems lowered wind speed and consequently soil erosion by wind (Sánchez and McCollin, 2015). Regarding the reduction of nitrate leaching, Nair et al. (2007) and Jose (2009) showed that agroforestry reduced nutrient losses by 40 to 70%. The conversion of 12 million ha of European cropland in NVZ to agroforestry with high tree densities could reduce nitrogen leaching by up to 28% (Palma et al., 2007). Moreno et al. (2016); Birrer et al. (2007); Bailey et al. (2010); and Lecq et al. (2017) investigated the potential of agroforestry to provide multiple habitats for flora and fauna and enhance biodiversity. Flowering trees, such as orchards with fruit trees, were especially important in providing nesting and foraging habitats for pollinators (Sutter et al., 2017) and could enhance pest control (Simon et al., 2011). In general, findings from recent literature suggest that green infrastructure, such as agroforestry, enhances the overall provision of multiple ecosystem services (Kay et al., 2018a,b; Maes et al., 2015).

#### 4.3. Carbon sequestration potential

Our third research question focussed on the most prominent pressure “climate change” in pursuit of a zero-emission scenario in European agriculture. To do this, we estimated the carbon storage potential of the proposed agroforestry systems in the above- and below-ground biomass of the woody elements. Whilst we are aware that agroforestry can also increase soil organic carbon (e.g. López-Díaz et al., 2017; Seitz et al., 2017; Upson and Burgess, 2013), soil carbon storage is difficult to quantify. E.g. Feliciano et al. (2018) reported inconsistent results for temperate agroforestry ranging from a decrease of  $-8 \text{ t C ha}^{-1} \text{ a}^{-1}$  to an increase of  $8 \text{ t C ha}^{-1} \text{ a}^{-1}$ . They affirmed that different climatic conditions and the previous land management had a higher impact on soil carbon storage than the established agroforestry system. At the scale of this study it was therefore not sufficiently reliable to account for (additional) soil carbon storage.

We found an overall average carbon sequestration potential of agroforestry of between  $0.09$  to  $7.29 \text{ t C ha}^{-1} \text{ a}^{-1}$ . The lower values were related to systems involving fewer woody elements per area (e.g. hedgerows on field boundaries, which typically make up less than 5% of the field). The higher values were mainly related to systems with higher densities of fast growing tree species and good soil conditions, which would also be associated with some reduction in food and feed production (see also Table 3). Previous studies (e.g. Palma et al., 2007; Reisner et al., 2007) estimated a sequestration range of between  $0.77$  and  $3 \text{ t C ha}^{-1} \text{ a}^{-1}$  for alley cropping, and Aertsens et al. (2013) proposed an average sequestration of  $2.75 \text{ t C ha}^{-1} \text{ a}^{-1}$ . Our estimates ranged from  $0.09$  to  $7.29 \text{ t C ha}^{-1} \text{ a}^{-1}$  for implementing different agroforestry systems across Europe. In comparison, European forest stands sequestered 167 million t C in 2015 on 160.93 million ha ( $1.04 \text{ t C ha}^{-1} \text{ a}^{-1}$ ) (FOREST EUROPE, 2015). This value is a continental average and also comprises trees grown at latitudes and altitudes where growth is relatively slow. In

<sup>1</sup> See also European Agroforestry Federation (EURAF) - <http://www.eurafagroforestry.eu/>

general, the competition between trees, e.g. for light and nutrients, is higher in forests than for trees in agroforestry systems.

#### 4.4. Potential implementation and impact

The hotspots of environmental pressures were mainly located, as was expected, in intensively managed agricultural regions mostly correlated with a high level of production (Eurostat, 2018, 2017c). The implementation of agroforestry in these regions would have the greatest environmental benefits (Weissteiner et al., 2016). In spite of the rising awareness of the importance of improving the environment and the investment in supporting measures of the European and national Rural Development Programs of the EU Member States (Santiago-Freijanes et al., 2018), the impact on green infrastructure is mixed. For example in the UK, whilst the area of woodland is increasing, the area of hedgerows declined from 1998 to 2007 (Wood et al., 2018). Agroforestry, landscape features, agro-ecological systems, and green infrastructure are still in decline (Angelstam et al., 2017; EEA, 2018; Salomaa et al., 2017). This implies that the established incentives are insufficient or do not adequately address the problem and actors (e.g. Mosquera-Losada et al., 2018). In contrast, a promising trend can be observed in Switzerland, where since 1993 agroforestry trees and hedgerows in open landscapes are qualified as ecological focus areas. This measure and the related payments have allowed a consolidation of the area under agroforestry (BLW, 2017; Herzog et al., 2018).

There might be a trade-off between the introduction of agroforestry on arable and grassland, food production and the challenge of food security over the coming decades with a rising human population (Ray et al., 2013). For example, for a poplar silvoarable system in the UK, García de Jalón et al. (2018) predicted that crop yields would be 42% of those in arable systems, and that timber yields would be 85% of those in a widely-spaced forest system. Thus, the crop production and hence the production of food for human nutrition would be reduced. In the case of silvopastoral practices, Rivest et al. (2013) showed that trees did not compromise pasture yields, though the impact of future drought pressures on yield would strongly be related to the chosen species. In addition, no significant correlation between the number of semi-natural vegetation on agricultural output was found (García-Feced et al., 2015).

The potential reduction of agricultural yields after the introduction of trees is an argument that is often put forward by farmers, who see themselves foremost as producers of food and fodder. However, under Mediterranean conditions, Arenas-Corraliza et al. (2018) predict that crop production could be reinforced under silvoarable schemes compared to open fields if the recurrence of warm springs keeps increasing. In addition, farmers are increasingly being asked to provide environmental goods and services beyond food production and policy makers and researchers are seeking for ways to sustainably intensify agricultural production, which necessitates increasing productivity whilst at the same time reducing environmental damage and maintaining the functioning of agro-ecosystems in the long-term (Tilman et al., 2011; Tilman and Clark, 2014). In many cases, this will require a shift towards more complex and knowledge intensive agro-ecological approaches (Garibaldi et al., 2017). Trees on farmland have been identified for a long time as key elements in the design of sustainable agricultural systems (Edwards et al., 1993) and can contribute to multiple ecosystem services beyond carbon sequestration in combination with other types of semi-natural vegetation (Smith et al., 2017).

Agroforestry implementation in the Priority Areas, which made up 8.9% of total European farmland, would capture between 1.4 and 43.4% of European agricultural GHG emissions, depending on whether the focus is on increasing tree cover in hedgerows as field boundary or supporting within field silvoarable and silvopastoral systems. These values support the observation by Hart et al. (2017) and Aertsens et al. (2013) who championed agroforestry as the most promising tool for climate change mitigation and adaptation in agriculture. Consequently, agroforestry can contribute significantly to the ambitious climate

targets of the EU for a zero-emission agriculture.

Finally, implications of this study are not restricted to the agricultural sector. Promoting agroforestry should be part of a more general land use policy aiming at the design of multifunctional agricultural landscapes. Scholars maintain that this will require coordinated actions at scales larger than individual farms and suggest that mechanisms for coordination and integration between spatial planning and agricultural measures will need to be put in place (e.g. Landis, 2017; Rega, 2014). This is also in line with the European Biodiversity Strategy and the Communication on Green Infrastructure (COM(2013) 249 fin. 1), which advocates for the integration between green infrastructure and spatial planning to achieve the Strategy's Target 2 objectives – ecosystem services enhancement and ecosystem restoration. In this frame, agroforestry should be considered as a key component of green infrastructure and, in turn, green infrastructure can offer a suitable policy frame, beyond the CAP, to promote agroforestry.

## 5. Conclusion

We investigated the potential for implementing agroforestry in agricultural areas subject to multiple environmental pressures of agricultural land in Europe and its contribution to European climate and GHG emission reduction targets. We found around one quarter of European arable and pasture land to be affected by none or only one of nine analysed environmental pressures and not primarily in need of restoration through introduction of agroforestry. Pastures were less affected than arable lands. For the Pressure Areas, we propose a wide range of agroforestry practices, which could mitigate the environmental pressures. The collection confirms the huge potential of agroforestry (1) to be introduced and established in nearly every region in Europe and (2) to adapt to various contexts, ideas, and needs of farmers. The estimated potential carbon storage depends on the selected agroforestry practice. The evidence from this study, that agroforestry on 8.9% of European agricultural land could potentially store between 1.4 up to 43.4% of the total European agricultural GHG emissions, is encouraging and demonstrates that agroforestry could contribute strongly to prepare the ground for future zero-emission agriculture. Imposing e.g. carbon payments or penalties for nutrient or soil loss pollutions as presented would make agroforestry a more financially profitable system. Future analysis should regionalize the approach to individual countries making use of data of higher spatial and thematic resolution, and ultimately to the farm scale, accompanied by extension and advice.

In sum, agroforestry can play a major role to reach national, European and global climate targets, whilst additionally fostering environmental policy and promoting sustainable agriculture, particularly in areas of intensive agricultural management where environmental pressures accumulate. Future policy and legislation, e.g. the future Common Agricultural Policy (CAP2020+), should explicitly promote and strengthen agroforestry.

## Acknowledgements

We acknowledge funding through Grant 613520 from the European Commission (Project AGFORWARD, 7th Framework Program). JVRD was supported by the Government of Asturias and the FP7-Marie Curie-COFUND program of the European Commission (Grant 'Clarín' ACA17-02). We would like to thank three anonymous reviewers for their constructive comments, which helped to improve an earlier version of the manuscript.

## Appendix A. Supplementary data

Supplementary data associated with this article can be found, in the online version, at <https://doi.org/10.1016/j.landusepol.2019.02.025>

## References

- Aertsens, J., Nocker, L., De, Gobin, A., 2013. Valuing the carbon sequestration potential for European agriculture. *Land Use Policy* 31, 584–594. <https://doi.org/10.1016/j.landusepol.2012.09.003>.
- Alam, M., Olivier, A., Paquette, A., Dupras, J., Revéret, J.P., Messier, C., 2014. A general framework for the quantification and valuation of ecosystem services of tree-based intercropping systems. *Agrofor. Syst.* 88, 679–691. <https://doi.org/10.1007/s10457-014-9681-x>.
- Angelstam, P., Khaulyak, O., Yamelynets, T., Mozgeris, G., Naumov, V., Chmielewski, T.J., Elbakidze, M., Manton, M., Prots, B., Valasiuk, S., 2017. Green infrastructure development at European Union's eastern border: effects of road infrastructure and forest habitat loss. *J. Environ. Manage.* <https://doi.org/10.1016/j.jenvman.2017.02.017>.
- Arenas-Corraliza, M.G., López-Díaz, M.L., Moreno, G., 2018. Winter cereal production in a Mediterranean silvoarable walnut system in the face of climate change. *Agric. Ecosyst. Environ.* 264, 111–118. <https://doi.org/10.1016/j.agee.2018.05.024>.
- BAFU, 2015. Diffuse Stickstoffeinträge in die Gewässer, modellierte Werte. [WWW Document]. URL <https://www.bafu.admin.ch/bafu/de/home/zustand/daten/geodaten/wasser-geodaten.html> (accessed 1.1.17).
- Bailey, D., Schmidt-Entling, M.H., Eberhart, P., Herrmann, J.D., Hofer, G., Kormann, U., Herzog, F., 2010. Effects of habitat amount and isolation on biodiversity in fragmented traditional orchards. *J. Appl. Ecol.* 47, 1003–1013. <https://doi.org/10.1111/j.1365-2664.2010.01858.x>.
- Birrer, S., Spiess, M., Herzog, F., Jenny, M., Kohli, L., Lugin, B., 2007. The Swiss agri-environment scheme promotes farmland birds: but only moderately. *J. Ornithol.* 148. <https://doi.org/10.1007/s10336-007-0237-y>.
- BLW, 2017. Agrarbericht 2017. Bundesamt für Landwirtschaft 460 p.
- BMUB, 2017. Stickstoffeintrag in die Biosphäre - Erster Stickstoff-Bericht der Bundesregierung. Bundesministerium für Umwelt, Naturschutz, Bau und Reakt, pp. 32.
- Borrelli, P., Lugato, E., Montanarella, L., Panagos, P., 2017. A new assessment of soil loss due to wind erosion in European agricultural soils using a quantitative spatially distributed modelling approach. *L. Degrad. Dev.* 28, 335–344. <https://doi.org/10.1002/ldr.2588>.
- Ceballos, A., Schnabel, S., 1998. Hydrological behaviour of a small catchment in the dehesa landuse system (Extremadura, SW Spain). *J. Hydrol.* 210, 146–160. [https://doi.org/10.1016/S0022-1694\(98\)00180-2](https://doi.org/10.1016/S0022-1694(98)00180-2).
- den Herder, M., Moreno, G., Mosquera-Losada, R.M., Palma, J.H.N., Sidiropoulou, A., Santiago Freijanes, J.J., Crous-Duran, J., Paulo, J.A., Tomé, M., Pantera, A., Papanastasis, V.P., Mantzanas, K., Pachana, P., Papadopoulos, A., Plieninger, T., Burgess, P.J., 2017. Current extent and stratification of agroforestry in the European Union. *Agric. Ecosyst. Environ.* 241, 121–132. <https://doi.org/10.1016/j.agee.2017.03.005>.
- DG Agriculture, Rural Development, 2017. Risk Management Schemes in EU Agriculture. *EU Agric. Mark. Briefs*. pp. 16.
- Edwards, C.A., Grove, T.L., Harwood, R.R., Pierce Colfer, C.J., 1993. The role of agroecology and integrated farming systems in agricultural sustainability. *Agric. Ecosyst. Environ.* 46, 99–121. [https://doi.org/10.1016/0167-8809\(93\)90017-J](https://doi.org/10.1016/0167-8809(93)90017-J).
- EEA, 2015a. Natura 2000 Data - the European Network of Protected Sites. [WWW Document]. Spat. dataset. URL <https://www.eea.europa.eu/data-and-maps/data/natura-9>.
- EEA, 2015b. High Nature Value (HNV) Farmland. [WWW Document]. Spat. dataset. URL <https://www.eea.europa.eu/data-and-maps/data/high-nature-value-farmland>.
- EEA, 2016. Corine Land Cover (CLC) 2012, Version 18.5.1. European Environment Agency (EEA) under the framework of the Copernicus programme - copernicus@eea.europa.eu.
- EEA, 2017a. Agriculture. [WWW Document]. Homepage. URL <https://www.eea.europa.eu/data-and-maps/data/7f36a220464ba44c85740244f15bc22e> (accessed 3.22.18).
- EEA, 2017b. Climate Change Adaptation and Disaster Risk Reduction in Europe. Enhancing coherence of the knowledge base, policies and practices - European Environment Agency, EEA Report <https://doi.org/10.1007/s13398-014-0173-7.2>.
- EEA, 2018. Report on Implementation Measures (Article 17, Habitats Directive) - Chart — Conservation Status and Trends of Habitats Assessed As Unfavourable, Per Annex I Category. [WWW Document]. URL <https://www.eea.europa.eu/data-and-maps/data/78a8fd22fa14fddb8ff218071aeb5d8> (accessed 1.4.18).
- ESRI, 2016. ArcGIS Desktop: Release 10.4. Environmental Systems Research Institute, Redlands, CA.
- European Commission, 2013a. Regulation (EU) No 1305/2013 on support for rural development by the European Agricultural Fund for Rural Development (EAFRD) and repealing Council Regulation (EC) No 1698/2005. *Off. J. Eur. Union* 1698, 487–548.
- European Commission, 2013b. Nitrogen Pollution and the European Environment: Implications for Air Quality Policy. *Sci. Environ. Policy IN-DEPTH Rep.* 28.
- European Commission, 2014. COMMISSION IMPLEMENTING REGULATION (EU) No 834/2014 of 22 July 2014. EU.
- European Commission, 2016. Agriculture - A Partnership between Europe and Farmers. The EU's Common Agricultural Policy (CAP): for our food, for our Countryside, for our Environment.
- European Commission, 2017a. Agriculture and the environment: Introduction. [WWW Document]. URL [https://ec.europa.eu/agriculture/envir\\_en](https://ec.europa.eu/agriculture/envir_en).
- European Commission, 2017b. Context Indicators (last update December 2017).
- Eurostat, 2017a. Share of Irrigated Land on Total Agricultural Area. [WWW Document]. *Farm Struct. Surv.* 2013.
- Eurostat, 2017b. Greenhouse Gas Emission Statistics - Emission Inventories. [WWW Document]. Stat. Explain. URL [http://ec.europa.eu/eurostat/statistics-explained/index.php/Greenhouse\\_gas\\_emission\\_statistics\\_-\\_emission\\_inventories](http://ec.europa.eu/eurostat/statistics-explained/index.php/Greenhouse_gas_emission_statistics_-_emission_inventories).
- Eurostat, 2017c. Archive: Agri-environmental Indicator - Soil Quality. [WWW Document]. Stat. Explain. URL [http://ec.europa.eu/eurostat/statistics-explained/index.php/Archive:Agri-environmental\\_indicator\\_-\\_soil\\_quality](http://ec.europa.eu/eurostat/statistics-explained/index.php/Archive:Agri-environmental_indicator_-_soil_quality).
- Eurostat, 2018. Agri-Environmental Indicator - Cropping Patterns. [WWW Document]. Stat. Explain. URL [http://ec.europa.eu/eurostat/statistics-explained/index.php/Agri-environmental\\_indicator\\_-\\_cropping\\_patterns#Agri-environmental\\_context](http://ec.europa.eu/eurostat/statistics-explained/index.php/Agri-environmental_indicator_-_cropping_patterns#Agri-environmental_context).
- Feliciano, D., Ledo, A., Hillier, J., Nayak, D.R., 2018. Which agroforestry options give the greatest soil and above ground carbon benefits in different world regions? *Agric. Ecosyst. Environ.* 254, 117–129. <https://doi.org/10.1016/j.agee.2017.11.032>.
- FOREST EUROPE, 2015. State of Europe's forests 2015. *Minist. Conf. Prot. For. Eur.* 314 p.
- García de Jalón, S., Graves, A., Palma, J.H., Williams, A., Upson, M., Burgess, P.J., 2018. Modelling and valuing the environmental impacts of arable, forestry and agroforestry systems: a case study. *Agrofor. Syst.* 92, 1059–1073. <https://doi.org/10.1007/s10457-017-0128-z>.
- García-Feced, C., Weissteiner, C.J., Baraldi, A., Paracchini, M.L., Maes, J., Zulian, G., Kempen, M., Elbersen, B., Pérez-Soba, M., 2015. Semi-natural vegetation in agricultural land: European map and links to ecosystem service supply. *Agron. Sustain. Dev.* 35, 273–283. <https://doi.org/10.1007/s13593-014-0238-1>.
- Garibaldi, L.A., Gemmill-Herren, B., D'Annolfo, R., Graeb, B.E., Cunningham, S.A., Breeze, T.D., 2017. Farming approaches for greater biodiversity, livelihoods, and food security. *Trends Ecol. Evol.* 32, 68–80. <https://doi.org/10.1016/j.tree.2016.10.001>.
- Harrell, F.E., 2018. R Package "Hmisc" Version 4.1.1.
- Hart, K., Bartel, A., Menadue, H., Sedy, K., Fredrik-Larsen, A., Hjerpe, P., 2012. Methodologies for Climate Proofing Investments and Measures under Cohesion and Regional Policy and the Common Agricultural Policy: Identifying the climate risks related to rural areas and adaptation options. *Rep. DG Clim.* 22 p.
- Hart, K., Allen, B., Keenleyside, C., Nanni, S., Maréchal, A., Paquel, K., Nesbit, M., Ziemann, J., 2017. Research for agri committee - the consequences of climate change for EU agriculture. Follow-Up To the COP21 - Un Paris Climate Change Conference. <https://doi.org/10.2861/295025>.
- Herzog, F., Szerencsits, E., Kay, S., Rocas-Díaz, J.V., Jäger, M., 2018. Agroforestry in Switzerland - a non-CAP European Country. *Agroforestry as Sustainable Land Use, in: 4th European Agroforestry Conference* 74–78.
- Hijmans, R.J., Cameron, S.E., Parra, J.L., Jones, P.G., Jarvis, A., 2005. Very high resolution interpolated climate surfaces for global land areas. *Int. J. Climatol.* 25, 1965–1978. <https://doi.org/10.1002/joc.1276>.
- Jose, S., 2009. Agroforestry for ecosystem services and environmental benefits: an overview. *Agric. Syst.* 76, 1–10. <https://doi.org/10.1007/s10457-009-9229-7>.
- JRC, 2017. Crop Monitoring in Europe - July 2017 25. *JRC MARS Bull.* pp. 42.
- JRC Water Portal, 2017. European Irrigation Map and the Assessment of Irrigation Requirements Aggregated to a Grid of 10x10km. Dataset of Irrigated Areas (1/0). [WWW Document]. Dataset. URL <http://water.jrc.ec.europa.eu/waterportal>.
- Kay, S., Crous-Duran, J., García de Jalón, S., Graves, A., Palma, J., Rocas-Díaz, J.V., Szerencsits, E., Weibel, R., Herzog, F., 2018a. Landscape-scale modelling of agroforestry ecosystem services in Swiss orchards: a methodological approach. *Landsc. Ecol.* 33, 1633–1644. <https://doi.org/10.1007/s10980-018-0691>.
- Kay, S., Crous-Duran, J., García de Jalón, S., Graves, A., Ferreiro-Domínguez, N., Moreno, G., Mosquera-Losada, M.R., Palma, J.H., Rocas-Díaz, J.V., Santiago-Freijanes, J.J., Szerencsits, E., Weibel, R., Herzog, F., 2018b. Spatial similarities between European agroforestry systems and ecosystem services at the landscape scale. *Agrofor. Syst.* 92, 1075–1089. <https://doi.org/10.1007/s10457-017-0132-3>.
- Kim, D.-G., Kirschbaum, M.U.F., Beedy, T.L., 2016. Carbon sequestration and net emissions of CH4 and N2O under agroforestry: synthesizing available data and suggestions for future studies. *Agric. Ecosyst. Environ.* 226, 65–78. <https://doi.org/10.1016/j.agee.2016.04.011>.
- Knox, J., Daccache, A., Hess, T., Haro, D., 2016. Meta-analysis of climate impacts and uncertainty on crop yields in Europe. *Environ. Res. Lett.* 11. <https://doi.org/10.1088/1748-9326/7/3/034032>.
- Landis, D.A., 2017. Designing agricultural landscapes for biodiversity-based ecosystem services. *Basic Appl. Ecol.* 18, 1–12. <https://doi.org/10.1016/j.baae.2016.07.005>.
- Lecq, S., Loisel, A., Brischoux, F., Mullin, S.J., Bonnet, X., 2017. Importance of ground refuges for the biodiversity in agricultural hedgerows. *Ecol. Indic.* 72, 615–626. <https://doi.org/10.1016/j.ecolind.2016.08.032>.
- Leip, A., Weiss, F., Lesschen, J.P., Westhoek, H., 2014. The nitrogen footprint of food products in the European Union. *J. Agric. Sci.* 152, 20–33. <https://doi.org/10.1017/S0021859613000786>.
- López-Díaz, M.L., Benítez, R., Moreno, G., 2017. How management techniques affect carbon sequestration in intensive hardwood plantations? *For. Ecol. Manage.* 389, 228–239. <https://doi.org/10.1016/j.foreco.2016.11.048>.
- Lugato, E., Bampa, F., Panagos, P., Montanarella, L., Jones, A., 2014a. Potential carbon sequestration of European arable soils estimated by modelling a comprehensive set of management practices. *Glob. Chang. Biol.* 20, 3557–3567. <https://doi.org/10.1111/gcb.12551>.
- Lugato, E., Panagos, P., Bampa, F., Jones, A., Montanarella, L., 2014b. A new baseline of organic carbon stock in European agricultural soils using a modelling approach. *Glob. Chang. Biol.* 20, 313–326. <https://doi.org/10.1111/gcb.12292>.
- Maes, J., Barbosa, A., Baranzelli, C., Zulian, G., Batista e Silva, F., Vandecasteele, I., Hiederer, R., Liqueite, C., Paracchini, M.L., Mubareka, S., Jacobs-Crisioni, C., Castillo, C.P., Laval, C., 2015. More green infrastructure is required to maintain ecosystem services under current trends in land-use change in Europe. *Landsc. Ecol.* 30, 517–534. <https://doi.org/10.1007/s10980-014-0083-2>.
- Malek, Z., Verbürg, P., 2017. Mediterranean land systems: representing diversity and intensity of complex land systems in a dynamic region. *Landsc. Urban Plan.* <https://doi.org/10.1016/j.landurbplan.2017.05.011>.

- [doi.org/10.1016/j.landurbplan.2017.05.012](https://doi.org/10.1016/j.landurbplan.2017.05.012).
- Martin, G.M., Bellouin, N., Collins, W.J., Culverwell, I.D., Halloran, P.R., Hardiman, S.C., Hinton, T.J., Jones, C.D., McDonald, R.E., McLaren, A.J., O'Connor, F.M., Roberts, M.J., Rodriguez, J.M., Woodward, S., Best, M.J., Brooks, M.E., Brown, A.R., Butchart, N., Dearden, C., Derbyshire, S.H., Dharssi, I., Doutriaux-Boucher, M., Edwards, J.M., Falloon, P.D., Gedney, N., Gray, L.J., Hewitt, H.T., Hobson, M., Huddleston, M.R., Hughes, J., Ineson, S., Ingram, W.J., James, P.M., Johns, T.C., Johnson, C.E., Jones, A., Jones, C.P., Joshi, M.M., Keen, A.B., Liddicoat, S., Lock, A.P., Maidens, A.V., Manners, J.C., Milton, S.F., Rae, J.G.L., Ridley, J.K., Sellar, A., Senior, C.A., Totterdell, I.J., Verhoef, A., Vidale, P.L., Wiltshire, A., 2011. The HadGEM2 family of met office unified model climate configurations. *Geosci. Model. Dev.* <https://doi.org/10.5194/gmd-4-723-2011>.
- McIvor, I., Youjun, H., Daoping, L., Eyles, G., Pu, Z., 2014. Agroforestry: conservation trees and erosion prevention. *Encycl. Agric. Food Syst.* 1, 208–221. <https://doi.org/10.1016/B978-0-444-52512-3.00247-3>.
- Moreno, G., Arenas, G., Juarez, E., 2016. System Report: Cereal Production Beneath Walnut in Spain.
- Mosquera-Losada, M.R., Santiago-Freijanes, J.J., Rois-Díaz, M., Moreno, G., den Herder, M., Aldrey-Vázquez, J.A., Ferreiro-Domínguez, N., Pantera, A., Pisanelli, A., Rigueiro-Rodríguez, A., 2018. Agroforestry in Europe: A land management policy tool to combat climate change. *Land Use Policy* 78, 603–613. <https://doi.org/10.1016/J.LANDUSEPOL.2018.06.052>.
- Mouchet, M.A., Paracchini, M.L., Schulp, C.J.E., Stürck, J., Verkerk, P.J., Verburg, P.H., Lavorel, S., 2017. Bundles of ecosystem (dis)services and multifunctionality across European landscapes. *Ecol. Indic.* 73, 23–28. <https://doi.org/10.1016/j.ecolind.2016.09.026>.
- Mücher, C.A., Klijn, J.A., Wascher, D.M., Schaminée, J.H.J., 2010. A new European Landscape Classification (LANMAP): a transparent, flexible and user-oriented methodology to distinguish landscapes. *Ecol. Indic.* 10, 87–103. <https://doi.org/10.1016/j.ecolind.2009.03.018>.
- Murphy, B.W., 2015. Impact of soil organic matter on soil properties - A review with emphasis on Australian soils. *Soil Res.* 53, 605–635. <https://doi.org/10.1071/SR14246>.
- Nair, V.D., Nair, P.K.R., Kalmbacher, R.S., Ezenwa, I.V., 2007. Reducing nutrient loss from farms through silvopastoral practices in coarse-textured soils of Florida, (USA). *Ecol. Eng.* 29, 192–199. <https://doi.org/10.1016/j.ecoleng.2006.07.003>.
- Olesen, J.E., Børgesen, C.D., Elsgaard, L., Palosuo, T., Rötter, R.P., Skjelvåg, A.O., Peltonen-Sainio, P., Börjesson, T., Trnka, M., Ewert, F., Siebert, S., Brissun, N., Eitzinger, J., van Asselt, E.D., Oberforster, M., van der Fels-Klerx, H.J., 2012. Changes in time of sowing, flowering and maturity of cereals in Europe under climate change. *Food Addit. Contam. Part A* 29, 1527–1542. <https://doi.org/10.1080/19440049.2012.712060>.
- Orgiazzi, A., Panagos, P., Yigini, Y., Dunbar, M.B., Gardi, C., Montanarella, L., Ballabio, C., 2016. A knowledge-based approach to estimating the magnitude and spatial patterns of potential threats to soil biodiversity. *Sci. Total Environ.* 545–546, 11–20. <https://doi.org/10.1016/j.scitotenv.2015.12.092>.
- Palma, J.H.N., Graves, A.R., Bunce, R.G.H., Burgess, P.J., de Filippi, R., Keesman, K.J., van Keulen, H., Liagre, F., Mayus, M., Moreno, G., Reisner, Y., Herzog, F., 2007. Modelling environmental benefits of silvoarable agroforestry in Europe. *Agric. Ecosyst. Environ.* 119, 320–334.
- Panagos, P., Borrelli, P., Poesen, J., Ballabio, C., Lugato, E., Meusburger, K., Montanarella, L., Alewell, C., 2015. The new assessment of soil loss by water erosion in Europe. *Environ. Sci. Policy* 54, 438–447. <https://doi.org/10.1016/j.envsci.2015.08.012>.
- Paracchini, M.L., Petersen, J.-E., Hoogeveen, Y., Bamps, C., Burfield, I., Swaay, C., Van, 2008. High Nature Value Farmland in Europe - An Estimate of the Distribution Patterns on the Basis of Land Cover and Biodiversity Data. Institute for Environment and Sustainability Office for Official Publications of the European Communities Luxembourg <https://doi.org/10.2788/8891>.
- Plieninger, T., Draux, H., Fagerholm, N., Bieling, C., Bürgi, M., Kizos, T., Kuemmerle, T., Primdahl, J., Verburg, P.H., 2016. The driving forces of landscape change in Europe: a systematic review of the evidence. *Land Use Policy* 57, 204–214. <https://doi.org/10.1016/j.landusepol.2016.04.040>.
- Prasuhn, V., Liniger, H., Gisler, S., Herweg, K., Candinas, A., Clément, J.-P.P., 2013. A high-resolution soil erosion risk map of Switzerland as strategic policy support system. *Land Use Policy* 32, 281–291. <https://doi.org/10.1016/j.landusepol.2012.11.006>.
- R Development Core Team, 2016. R Software. R: A Language and Environment for Statistical Computing.
- Ray, D.K., Mueller, N.D., West, P.C., Foley, J.A., 2013. Yield trends are insufficient to double global crop production by 2050. *PLoS One* 8. <https://doi.org/10.1371/journal.pone.0066428>.
- Rega, C., 2014. Pursuing Integration Between Rural Development Policies and Landscape Planning: Towards a Territorial Governance Approach. SpringerBriefs Geogr.
- Rega, C., Paracchini, M.L., Zulian, G., 2017. Quantification of Ecological Services for Sustainable Agriculture. QUESA Deliverable D4.4: Report on spatially explicit “heat maps” for ES across Europe, Ispra.
- Rega, C., Bartual, A.M., Bocci, G., Sutter, L., Albrecht, M., Moonen, A.-C., Jeanneret, P., van der Werf, W., Pfister, S.C., Holland, J.M., Paracchini, M.L., 2018. A pan-European model of landscape potential to support natural pest control services. *Ecol. Indic.* 90, 653–664. <https://doi.org/10.1016/j.ecolind.2018.03.075>.
- Reisner, Y., de Filippi, R., Herzog, F., Palma, J., 2007. Target regions for silvoarable agroforestry in Europe. *Ecol. Eng.* 29, 401–418. <https://doi.org/10.1016/j.ecoleng.2006.09.020>.
- Rivest, D., Paquette, A., Moreno, G., Messier, C., 2013. A meta-analysis reveals mostly neutral influence of scattered trees on pasture yield along with some contrasted effects depending on functional groups and rainfall conditions. *Agric. Ecosyst. Environ.* 165, 74–79.
- Rois-Díaz, M., Lovric, N., Lovric, M., Ferreiro-Domínguez, N., Mosquera-Losada, M.R., Den Herder, M., Graves, A., Palma, J.H.N., Paulo, J.A., Pisanelli, A., Smith, J., Moreno, G., García, S., Varga, A., Pantera, A., Mirck, J., Burgess, P., 2018. Farmers' reasoning behind the uptake of agroforestry practices: evidence from multiple case-studies across Europe. *Agrofor. Syst.* 92, 811–828. <https://doi.org/10.1007/s10457-017-0139-9>.
- Salomaa, A., Paloniemi, R., Kotiaho, J.S., Kettunen, M., Apostolopoulou, E., Cent, J., 2017. Can green infrastructure help to conserve biodiversity? *Environ. Plan. C Gov. Policy* 35, 265–288. <https://doi.org/10.1177/0263774X16649363>.
- Sánchez, I.A., McCollin, D., 2015. A comparison of microclimate and environmental modification produced by hedgerows and dehesa in the Mediterranean region: A study in the Guadarrama region, Spain. *Landsc. Urban Plan.* 143, 230–237. <https://doi.org/10.1016/j.landurbplan.2015.07.002>.
- Santiago-Freijanes, J.J., Rigueiro-Rodríguez, A., Aldrey-Vázquez, J.A., Moreno-Marcos, G., den Herder, M., Burgess, P.J., Mosquera-Losada, M.R., 2018. Understanding Agroforestry practices in Europe through landscape features policy promotion. *Agrofor. Syst.* 92, 1105–1115. <https://doi.org/10.1007/s10457-018-0212-z>.
- Schauberger, B., Archontoulis, S., Arnett, A., Balkovic, J., Ciaia, P., Deryng, D., Elliott, J., Folberth, C., Khabarov, N., Müller, C., Pugh, T.A.M., Rolinski, S., Schaphoff, S., Schmid, E., Wang, X., Schlenker, W., Frieler, K., 2017. Consistent negative response of US crops to high temperatures in observations and crop models. *Nat. Commun.* 8. <https://doi.org/10.1038/ncomms13931>.
- Schulp, Burkhard, B., Maes, J., Van Vliet, J., Verburg, P.H., 2014. Uncertainties in ecosystem service maps: a comparison on the European scale. *PLoS One* 9. <https://doi.org/10.1371/journal.pone.0109643>.
- Seitz, B., Carrand, E., Burgos, S., Tatti, D., Herzog, F., Jäger, M., Sereke, F., 2017. Erhöhte Humusvorräte in einem siebenjährigen Agroforstsystem in der Zentralschweiz / Augmentation des stocks d'humus dans un système agroforestier de sept ans en Suisse centrale. *Agrar. Schweiz* 8, 318–323.
- Simon, S., Bouvier, J.-C., Debras, J.-F., Sauphanor, B., 2011. Biodiversity and Pest Management in Orchard Systems. *Sustain. Agric* 2 (30), 693–709. [https://doi.org/10.1007/978-94-007-0394-0\\_30](https://doi.org/10.1007/978-94-007-0394-0_30).
- Smith, A.C., Harrison, P.A., Pérez Soba, M., Archaux, F., Blicharska, M., Egoh, B.N., Erős, T., Fabrega Domenech, N., György, I., Haines-Young, R., Li, S., Lommelen, E., Meiresonne, L., Miguel Ayala, L., Mononen, L., Simpson, G., Stange, E., Turkelboom, F., Uiterwijk, M., Veerkamp, C.J., Wyllie de Echeverria, V., 2017. How natural capital delivers ecosystem services: a typology derived from a systematic review. *Ecosyst. Serv.* 26, 111–126. <https://doi.org/10.1016/j.ecoser.2017.06.006>.
- Sutter, L., Herzog, F., Dietemann, V., Charrière, J.D., Albrecht, M., 2017. Nachfrage, Angebot und Wert der Insektenbestäubung in der Schweizer Landwirtschaft. *Agrar. Schweiz* 8, 332–339.
- Tilman, D., Clark, M., 2014. Global diets link environmental sustainability and human health. *Nature* 515, 518–522. <https://doi.org/10.1038/nature13959>.
- Tilman, D., Balzer, C., Hill, J., Befort, B.L., 2011. Global food demand and the sustainable intensification of agriculture. *Proc. Natl. Acad. Sci.* 108, 20260–20264. <https://doi.org/10.1073/pnas.1116437108>.
- Torralba, M., Fagerholm, N., Burgess, P.J., Moreno, G., Plieninger, T., 2016. Do European agroforestry systems enhance biodiversity and ecosystem services? A meta-analysis. *Agric. Ecosyst. Environ.* 230, 150–161. <https://doi.org/10.1016/j.agee.2016.06.002>.
- UNFCCC, 2015. Paris Agreement, Conference of the Parties on Its Twenty-First Session. doi:FCCC/CP/2015/L.9/Rev.1.
- Upon, M.A., Burgess, P.J., 2013. Soil organic carbon and root distribution in a temperate arable agroforestry system. *Plant Soil* 373, 43–58. <https://doi.org/10.1007/s11040-013-1733-x>.
- Verheijen, F.G.A., Jones, R.J.A., Rickson, R.J., Smith, C.J., 2009. Tolerable versus actual soil erosion rates in Europe. *Earth-Sci. Rev.* 94, 23–38. <https://doi.org/10.1016/j.earscirev.2009.02.003>.
- Weissteiner, C.J., García-Feced, C., Paracchini, M.L., 2016. A new view on EU agricultural landscapes: quantifying patchiness to assess farmland heterogeneity. *Ecol. Indic.* 61, 317–327. <https://doi.org/10.1016/j.ecolind.2015.09.032>.
- Wezel, A., Casagrande, M., Celette, F., Vian, J.F., Ferrer, A., Peigné, J., 2014. Agroecological practices for sustainable agriculture. A review. *Agron. Sustain. Dev.* 34, 1–20. <https://doi.org/10.1007/s13593-013-0180-7>.
- Wickham, H., 2016. plyr: Tools for Splitting, Applying and Combining Data.
- Wickham, H., Winston, C., RStudio, 2016. ggplot2: Create Elegant Data Visualisations Using the Grammar of Graphics. Cran. <https://doi.org/10.1093/bioinformatics/btr406>.
- Wood, C.M., Bunce, R.G.H., Norton, L.R., Maskell, L.C., Smart, S.M., Scott, W.A., Henrys, P.A., Howard, D.C., Wright, S.M., Brown, M.J., Scott, R.J., Stuart, R.C., Watkins, J.W., 2018. Ecological landscape elements: long-term monitoring in Great Britain, the Countryside Survey 1978–2007 and beyond. *Earth Syst. Sci. Data* 10, 745–763. <https://doi.org/10.5194/essd-10-745-2018>.
- Zhang, W., Ricketts, T.H., Kremen, C., Carney, K., Swinton, S.M., 2007. Ecosystem services and dis-services to agriculture. *Ecol. Econ.* 64, 253–260. <https://doi.org/10.1016/j.ecolecon.2007.02.024>.